

1 **Title**

2 Retene, pyrene and phenanthrene cause distinct molecular-level changes in the cardiac tissue of
3 rainbow trout (*Oncorhynchus mykiss*) larvae, Part 2 – Proteomics and metabolomics

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17 **Abstract**

18 Polycyclic aromatic hydrocarbons (PAHs) are global contaminants of concern. Despite several
19 decades of research, their mechanisms of toxicity are not very well understood. Early life stages
20 of fish are particularly sensitive with the developing cardiac tissue being a main target of PAHs

21 toxicity. The mechanisms of cardiotoxicity of the three widespread model polycyclic aromatic
22 hydrocarbons (PAHs) retene, pyrene and phenanthrene were explored in rainbow trout
23 (*Oncorhynchus mykiss*) early life stages. Newly hatched larvae were exposed to sublethal doses
24 of each individual PAH causing no detectable morphometric alterations. Changes in the cardiac
25 proteome and metabolome were assessed after 7 or 14 days of exposure to each PAH. Phase I
26 and II enzymes regulated by the aryl hydrocarbon receptor were significantly induced by all
27 PAHs, with retene being the most potent compound. Retene significantly altered the level of
28 several proteins involved in key cardiac functions such as muscle contraction, cellular tight
29 junctions or calcium homeostasis. Those findings were quite consistent with previous reports
30 regarding the effects of retene on the cardiac transcriptome. Significant changes in proteins
31 linked to iron and heme metabolism were observed following exposure to pyrene. While
32 phenanthrene also altered the levels of several proteins in the cardiac tissue, no clear mechanisms
33 or pathways could be highlighted. Due to high variability between samples, very few significant
34 changes were detected in the cardiac metabolome overall. Slight but significant changes were
35 still observed for pyrene and phenanthrene, suggesting possible effects on several energetic or
36 signaling pathways. This study shows that early exposure to different PAHs can alter the
37 expression of key proteins involved in the cardiac function, which could potentially affect
38 negatively the fitness of the larvae and later of the juvenile fish.

39 **Keywords:** aquatic toxicology, cardiotoxicity, developmental toxicity, metabolomics, polycyclic
40 aromatic hydrocarbons (PAHs), proteomics

41 **1. Introduction**

42 Polycyclic aromatic hydrocarbons (PAHs) are among the most ubiquitous environmental
43 contaminants of global concern (Rubailo and Oberenko, 2008). Originating from both pyrogenic

44 or petrogenic sources, their influx into to the environment is believed to have increased over the
45 past decades because of increasing human activities (Kurek et al., 2013). PAHs usually end up in
46 the aquatic environment following for example atmospheric deposition, human-related effluents
47 (municipal or industrial), surface runoffs or natural and accidental oil spills (Wolska et al., 2012).
48 Aquatic organisms are chronically exposed to PAHs by direct contact with the water column, the
49 sediment or by preying on other potentially contaminated organisms (Logan, 2007). During
50 accidental episodes (e.g. the 2010 *Deepwater Horizon* disaster) resulting in the release of large
51 quantities of oil in a short period of time into the aquatic environment, fish communities can be
52 exposed to concentrations of PAHs high enough to be threatening at the population level
53 (Incardona et al., 2014; Incardona et al., 2015). PAHs are listed as priority substances for the
54 aquatic environment under the EU Water Framework Directive (Water Framework Directive,
55 2000). In the US, the EPA has 16 PAHs classified as priority pollutants (Keith, 2015), an
56 approach that has sometimes being judged outdated and inadequate (Andersson and Achten,
57 2015).

58 Some PAHs are able to bind with the aryl hydrocarbon receptor (AhR) and induce dioxin-like
59 toxicity in early life stages (ELS) of fish (Billiard et al., 1999; Scott et al., 2011). Dioxin-like
60 toxicity in ELS of fish is often associated with the induction of detoxification enzymes (e.g.
61 cytochrome P4501A) and characterized by the so-called blue sac disease (BSD) syndrome,
62 which was first described in salmonids exposed to the model compound 2,3,7,8-
63 tetrachlorodibenzo-*p*-dioxin (TCDD). The BSD syndrome includes various symptoms such as
64 delayed growth, skull and jaw deformities, cardiovascular defects, pericardial and yolk sac
65 edemas, hemorrhages and potentially death. Most importantly, the cardiac tissue appears to be
66 the primary target of dioxin-like compounds (DLCs), as cardiovascular defects are the first

67 symptoms observed in fish ELS exposed to DLCs (Doering et al., 2019). However, not all PAHs
68 are able to bind efficiently with the AhR (Barron et al., 2004). Some PAHs described as weak
69 AhR agonists are able to induce AhR-independent cardiotoxicity in ELS of fish, including
70 symptoms such as pericardial edemas usually linked to DLCs (Incardona et al., 2005). Overall,
71 and despite almost two decades of research, the mechanisms of toxicity of PAHs are still poorly
72 understood.

73 Retene (7-isopropyl-1-methylphenantrene) is an alkylated three-ring PAH commonly found in
74 pulp and paper mill effluents (Leppänen and Oikari, 2001), but also formed by thermal
75 degradation of resin compounds during wood combustion (Shen et al., 2012). Retene was
76 reported at concentrations in the ng.L⁻¹ range in the surface water and in the ng.g⁻¹ range in the
77 sediment of North American lakes or rivers (Ahad et al., 2015; Geier et al., 2018; Ruge et al.,
78 2015). In sediments historically contaminated by pulp and paper mill effluents, concentrations
79 ranging from several hundreds and several thousands of µg.g⁻¹ can be found (Leppänen and
80 Oikari, 2001; Meriläinen et al., 2006). Contaminants sequestered in the sediment can be released
81 and made bioavailable to fish when the sediment is disturbed (Eggleton and Thomas, 2004).

82 Retene is a potent AhR agonist capable of inducing dioxin-like toxicity in ELS of fish, as well as
83 significant changes in the cardiac transcriptome of rainbow trout (*Oncorhynchus mykiss*) larvae at
84 sublethal doses (Billiard et al., 1999; Scott et al., 2011; Vehniäinen et al., 2016). Retene was also
85 recently described as a potential mediator of the cardiac function by altering the action potential
86 as well as the intensity of ionic currents in rainbow trout ventricular cardiomyocytes exposed *in*
87 *vitro* (Vehniäinen et al., 2019). Pyrene is another widespread PAH known to be an AhR agonist
88 able to induce ELS dioxin-like toxicity in fish, including alteration of the cardiac function
89 (Barjhoux et al., 2014; Hendon et al., 2008; Shi et al., 2012; Zhang, Y. et al., 2012). Pyrene is a

90 weaker AhR agonist compared to retene (Barron et al., 2004). While the knockdown of *cyp1a*
91 using morpholino oligonucleotides appeared inefficient to protect from the embryotoxic effects
92 of retene in zebrafish (*Danio rerio*) (Scott et al., 2011), an opposite result was observed for
93 pyrene in the same species (Incardona et al., 2005). This suggests distinct mechanisms of toxicity
94 between retene and pyrene following AhR activation, or distinct non-AhR mediated pathways
95 between these two PAHs.

96 Phenanthrene is a three-ring PAH, known to be among the most abundant PAH compound in the
97 air, in precipitation and in coastal and estuarine waters and sediments (Latimer and Zheng,
98 2003). Pyrene and phenanthrene can reach concentrations in the $\mu\text{g.L}^{-1}$ range in surface water
99 impacted by industrialized areas, and up to several hundreds of $\mu\text{g.L}^{-1}$ close to crude oil
100 exploitation (Anyakora et al., 2005; Maskaoui et al., 2002). Phenanthrene is a good example of
101 PAHs that have a very low affinity with the AhR (Barron et al., 2004) but are still able to
102 produce cardiotoxicity in fish ELS. The cardiotoxic effects of phenanthrene have been described
103 in several fish species and include defects in heart looping, pericardial edemas, changes in the
104 cardiac rhythm (bradycardia and arrhythmias), atrioventricular conduction blockage and
105 reduction of blood circulation (Cypher et al., 2017; Incardona et al., 2004; Incardona et al., 2005;
106 Mu et al., 2014; Sun, L. et al., 2015; Zhang, Y., Huang, Zuo et al., 2013; Zhang, Y., Huang,
107 Wang et al., 2013). One possible mechanism of cardiotoxicity of phenanthrene in fish involves
108 the alteration of the action potential and key ionic currents, as demonstrated *in vitro* in
109 cardiomyocytes of the rainbow trout, the Pacific bluefin tuna (*Thunnus orientalis*) and the Pacific
110 mackerel (*Scomber japonicus*) (Brette et al., 2017; Vehniäinen et al., 2019).

111 The main objective of the present study was to investigate the mechanisms of cardiotoxicity of
112 the PAHs retene, pyrene and phenanthrene during the early life stages of a model fish species,

113 the rainbow trout. OMICs methods such as transcriptomics, proteomics and metabolomics have
114 gained popularity in ecotoxicological studies, as they allow to study hundreds of molecular
115 signals at the same time (Gündel et al., 2012). The present work is part of a larger project in
116 which several OMICs tools were combined in an attempt to better understand the mechanisms of
117 cardiotoxicity of PAHs. The changes in the cardiac transcriptome, proteome and metabolome
118 following exposure to those model PAHs were assessed using a similar experimental setup. This
119 manuscript reports the proteomic and metabolomic data, while the transcriptomic data was
120 published separately (submitted manuscript).

121 **2. Materials and methods**

122 2.1. Chemicals

123 Pyrene (>98% purity), phenanthrene ($\geq 99.5\%$ purity) and dimethyl sulfoxide (DMSO,
124 anhydrous, $\geq 99.9\%$ purity) were all purchased from Sigma-Aldrich (St-Louis, MO, USA).
125 Retene ($\geq 98\%$ purity) was obtained from MP Biomedicals (Illkirch, France). Stock solutions of
126 each individual PAH were prepared by dissolving them in DMSO to reach concentrations of 3.2
127 mg.mL⁻¹ for retene and pyrene, and 10 mg.mL⁻¹ for phenanthrene.

128 2.2. Experimental design

129 Exposure tanks (N = 52) consisted of 1.5L Pyrex glass bowls filled with 1L of filtered lake water
130 (Lake Konnevesi, Konnevesi, Finland) and were prepared 24h before the start of the exposures.
131 The dissolved organic carbon (DOC) in the filtered lake water ranged between 7,0 and 8,2 mg.L⁻¹
132 ¹(unpublished results). The bowls were randomly split between four different treatments (N = 13
133 tanks per treatment): DMSO (0.001%), retene (RET, 32 $\mu\text{g.L}^{-1}$), pyrene (PYR, 32 $\mu\text{g.L}^{-1}$) and
134 phenanthrene (PHE, 100 $\mu\text{g.L}^{-1}$). The quantity of DMSO per tank was the same in all tanks (10

135 μL of pure DMSO or of the appropriate PAH stock solution into 1L of lake water). These
136 exposure concentrations were selected based on previous studies in the case of retene (Billiard et
137 al., 1999), and from preliminary experiments (unpublished results) for pyrene and phenanthrene.
138 The concentrations used in the present study were chosen to be sublethal and to cause
139 cardiotoxic effects such as pericardial edemas and arrhythmias. Rainbow trout (*Oncorhynchus*
140 *mykiss*) eyed embryos at 360 degree-days ($^{\circ}\text{D}$) of development were obtained from a local fish
141 farm (Hanka-Taimen, Hankasalmi, Finland). Healthy (i.e. with no visible deformity) and newly
142 hatched (<24h) embryos were randomly distributed into twelve tanks per treatment (N = 15
143 embryos per bowl, 720 embryos in total). Four tanks (one per treatment) were left without fish,
144 for PAH concentration measurements. The water in the exposure tanks was completely renewed
145 daily and fresh chemicals were added to ensure constant exposure to PAHs (i.e. semi-static
146 exposure). The water temperature was measured every day and pH, conductivity and dissolved
147 oxygen were monitored on a regular basis. The light:dark cycle was set on 16h:8h, and yellow
148 fluorescent tubes were used. Larval mortality was monitored daily.

149 The procedure described above was reproduced three times, for three separate experiments: the
150 larvae were sampled after either 7 (one experiment) or 14 days (two experiments, but one with
151 only 9 tanks per treatment, including one without fish) of exposure. During sampling, all larvae
152 were scored for signs of BSD (pericardial and yolk sac edemas, hemorrhages, craniofacial and
153 spinal deformities) and a severity index ranging from 0 to 1 was calculated for each tank
154 according to Villalobos et al. (2000), with some minor modifications described in Scott and
155 Hodson (2008). All individuals were quickly and cautiously dissected under a microscope. The
156 heart of each larva was isolated from the rest of the body using fine forceps (Dumont #5, Fine
157 Science Tools, Heidelberg, Germany). The hearts of the larvae from two or three different tanks

158 per treatment were pooled together, thus each treatment consisted of four samples containing 30
159 or 45 hearts each. The samples (N = 32) used for the proteomics assay described in the present
160 study were the same samples as the ones used for the transcriptomics assay after 7 and 14 days of
161 exposure (submitted manuscript). The samples (N = 16) used for the metabolomics assay
162 described in the present study were from the second 14 days exposure described above. All
163 samples (N= 4 per treatment and each time point, for both the proteomics and the metabolomics)
164 were immediately frozen in liquid nitrogen and then stored at -80°C until further analyses.

165 2.3. Measurements of PAH concentration in water

166 Depending on the duration of the experiment (7 or 14 days), water samples were collected after
167 1, 3, 7, 10 and 14 days of exposure, before the daily renewal of the exposure water.

168 Concentrations measured by SFS right after the daily renewal are very close to the nominal
169 values (Honkanen et al., 2020; unpublished results). Sample collection was performed by
170 pipetting 5 mL of exposure water from four randomly chosen tanks per treatment, as well as
171 from each tank lacking fish at the beginning of the experiment (one per treatment). Before
172 storage at 4°C, 5 mL of ethanol (99.5% purity) was added in each sample.

173 The measurements were performed by synchronous fluorescence spectroscopy (SFS) using a LS-
174 55 Fluorescence Spectrometer (PerkinElmer, Waltham, MA, USA), following the method and
175 parameters previously described (submitted manuscript). More details regarding the method are
176 available in the Supplementary file 1. Retene and phenanthrene water samples from the
177 metabolomics exposure were affected by storage issues. Consequently, the only original data
178 presented in the present manuscript are the pyrene concentrations and fluorescence spectrums
179 from the 14 days exposure linked to the metabolomics assay. The PAH concentration data in
180 water and carcasses of the larvae for proteomics assay exposures has been already presented

181 (submitted manuscript). Detailed methods and results for both measurements are accessible in
182 the Supplementary file 1.

183 2.4. Protein extraction and measurements

184 The samples (N = 4) used for the protein extraction were the same as the ones used for the RNA
185 extraction in the 7 and 14 days exposures presented in a different manuscript (submitted
186 manuscript). After using the TRI Reagent (Molecular Research Center, Cincinnati, OH, USA) to
187 isolate the aqueous phase containing the RNA, ethyl alcohol (99.5% purity) was added to the
188 remaining organic phase and interphase to pellet the DNA. DNA was discarded and proteins
189 were precipitated using isopropanol. After a series of successive washes using 0.3M guanidine
190 hydrochloride (in 95% ethyl alcohol) and ethyl alcohol (99.5% purity), protein resuspension was
191 performed using a solution of 8M urea and 2M thiourea in a 1M Tris-HCl buffer (pH 8.0).

192 Before the in-solution trypsin digestion, 15 µg of protein for each sample were reduced with
193 dithiothreitol (1h at 37°C) and alkylated by iodoacetamide (1h at room temperature). Urea was
194 diluted below 1M using 50 mM Tris-HCl. Trypsin was added in ratio 1:30 (w/w) and the
195 samples were digested for 16h at 37°C. Digested peptides were desalted with SepPak C18 96-
196 well plate (Waters, Milford, MA, USA) according to the instructions of the manufacturer,
197 evaporated to dryness with SpeedVac (Thermo Fisher Scientific) and dissolved in 0.1% formic
198 acid before MS analysis. Peptide concentrations were determined with NanoDrop™ (Thermo
199 Fisher Scientific) by measuring absorbance at 280 nm. Concentrations of all samples were
200 adjusted to 100 ng.µL⁻¹.

201 The LC-ESI-MS/MS analyses were performed on a nanoflow HPLC system (Easy-nLC1200,
202 Thermo Fisher Scientific) coupled to the Q Exactive HF mass spectrometer (Thermo Fisher

203 Scientific) equipped with a nano-electrospray ionization source. Peptides were first loaded on a
204 trapping column and subsequently separated inline on a 15 cm C18 column (75 μm x 15 cm,
205 ReproSil-Pur 5 μm 200 \AA C18-AQ, Dr. Maisch HPLC GmbH, Ammerbuch-Entringen,
206 Germany). The mobile phase consisted of water with 0.1% formic acid (solvent A) or
207 acetonitrile/water (80:20 (v/v)) with 0.1% formic acid (solvent B). A linear 120 min two steps
208 gradient was used to elute peptides (85 min from 5% to 28% B, followed by 35 min from 28% to
209 40% B and finally 5 min wash with 100% B). All samples were injected twice, as technical
210 replicates.

211 2.5. Metabolite extraction and processing

212 Alanine-d₄ was obtained from Isotec (Sigma-Aldrich company), benzoic acid-d₅ and glycerol-
213 d₈ were from Campro (Berlin, Germany), salicylic acid-¹³C from Icon (Schlächtern, Germany),
214 and alkane standard (C₇-C₄₀) from Supelco (Sigma-Aldrich company). N-methyl-N-
215 (trimethylsilyl) trifluoroacetamide with 1% of trimethylsilyl chlorosilane (MSTFA with 1% of
216 TMCS) was purchased from Thermo Fisher Scientific (Waltham, MA, USA). All other
217 chemicals were from Sigma-Aldrich.

218 Metabolites were analyzed from three to four samples from each treatment (N = 3-4).

219 Metabolites were extracted in two steps. First, cold methanol with 0.1% of formic acid (300 μL)
220 and internal standard solution (10 μL) (0.24 $\text{mg}\cdot\text{mL}^{-1}$ of alanine-d₄, 0.9 $\text{mg}\cdot\text{mL}^{-1}$ of benzoic acid-
221 d₅, 0.25 $\text{mg}\cdot\text{mL}^{-1}$ of glycerol-d₈, 0.08 $\text{mg}\cdot\text{mL}^{-1}$ of salicylic acid-¹³C, 0.38 $\text{mg}\cdot\text{mL}^{-1}$ of 4-
222 methylumbelliferone in 8:8:3 H₂O:MeOH:DMSO) were added to the samples. Samples were
223 then homogenized with a bead mill (5 mm stainless steel beads, 2 \times 15s, 20 Hz, Qiagen
224 TissueLyser). The metabolites were extracted with an Eppendorf Thermomixer for 15 min at 4°C
225 at 1400 rpm. After a short centrifugation (2 min, 10°C, 13500g), supernatant was transferred to a

226 new test tube. The second extraction step was performed with cold 80% aqueous methanol with
227 0.1% formic acid (300 μL). Samples were homogenized with a beadmill, and metabolites were
228 extracted for 5 min at 4°C at 1400 rpm. Samples were shortly centrifuged and the two
229 supernatants were combined. Aliquots (200 μL) were transferred to vials and dried in a vacuum
230 at 35°C for 40 min. Quality control samples (QC) were prepared by combining extracts from
231 each sample group and included in each GC-MS analysis batch. Vials were treated with nitrogen
232 gas before they were capped and stored overnight at -70°C.

233 Samples were taken out of the freezer and allowed to reach room temperature before the vials
234 were opened. Dichloromethane (50 μL) was added to each sample and dried in a vacuum for 5
235 min. Samples were derivatized with 50 μL of 20 $\text{mg}\cdot\text{mL}^{-1}$ methoxyamine hydrochloride in
236 pyridine (MAHC) at 37°C for 90 min under continuous shaking (150 rpm). Samples were
237 silylated with 70 μL of MSTFA with 1% of TMSC at 37°C for 60 min under continuous shaking
238 (150 rpm). Alkane series in hexane (5 μL ; C7-C40) was added to each sample as a retention time
239 standard. Hexane (100 μL) was added to each sample before GC-MS analysis. GC-MS analysis
240 was performed with an Agilent 6890N chromatography system coupled with a 5973N mass
241 spectrometer, and a 7683 autosampler and injector. The injector was operated with pulsed
242 splitless mode (1 μL) with a 30 psi pulse for 0.60 min and purge flow at 0.50 min. Injection
243 temperature was set to 260°C and the sample was injected in a deactivated gooseneck splitless
244 liner with glass wool. Helium flow to the column (30 m Rxi-5Sil MS, 0.25 mm ID, 0.25 μm film
245 thickness with 10 m Integra-Guard, Restek) was kept constant at 1 $\text{mL}\cdot\text{min}^{-1}$ and purge flow was
246 46 $\text{mL}\cdot\text{min}^{-1}$. MSD interface temperature was 280°C, MS source 230°C and quadrupole 150°C.
247 The oven temperature program was as follows: at 60°C for 3 min, 7°C min^{-1} ramp to 240°C,
248 10°C min^{-1} ramp to 330°C, 2 min at 330°C and post-run at 60°C for 6 min. Mass spectra were

249 collected with a scan range of 55-550 m/z with 2.94 scans s^{-1} and for metabolite annotation QC
250 sample was analyzed with a wider scan range of 55-700 m/z .

251 Deconvolution, component detection and quantification were performed with Metabolite
252 Detector (versions 2.06 beta and 2.2N) (Hiller et al., 2009) and AMDIS (version 2.66, NIST).
253 Metabolite content was calculated as the peak area of the metabolite normalized with the peak
254 area of the internal standard, benzoic- d_5 acid, and the dry weight of the sample. Metabolites were
255 annotated based on spectra and retention index matched to reference compounds, and databases:
256 the Golm Metabolome database (GMD) (Hummel et al., 2007) and the NIST Mass Spectral
257 database (version 2.2 Agilent Technologies).

258 2.6. Bioinformatics and statistical analyses

259 2.6.1. Proteomics and metabolomics

260 Raw LC-ESI-MS/MS files were processed through MaxQuant 1.6.2.3 (Cox and Mann, 2008) for
261 protein identification and label-free quantification (LFQ). The search engine Andromeda (Cox et
262 al., 2011), integrated to the MaxQuant environment, was used to search proteins against two
263 RefSeq databases during two separate MaxQuant runs: one for *O. mykiss* and one for the Atlantic
264 Salmon (*Salmo salar*). Trypsin/P was used as the digestion mode parameter, with a maximum of
265 two missed cleavages allowed. Variable modifications were set on N-terminal acetylation,
266 oxidation of methionine and deamidation of asparagines and glutamines. The “match between
267 runs” parameter of MaxQuant was set on 0.7 min for the match time window and 20 min for the
268 alignment time window.

269 Statistical analyses were performed using R 3.5.1 (The R Foundation for Statistical Computing)
270 and Bioconductor 3.8, with the significant level set at $\alpha = 0.05$. The analyses of the MaxQuant

271 output were performed using the R package DEP 1.5.1 (Zhang, X. et al., 2018). This package
272 was favorited over the traditionally used Perseus software in order to apply the LIMMA
273 approach (Linear Models for Microarray Data) to our dataset. This approach has been proved
274 more powerful than the ordinary *t*-tests when the number of samples or replicates is small
275 (Kammers et al., 2015). Proteins detected with less than two peptides were removed from the
276 dataset. The data was filtered for proteins which were not quantified in all replicates: only the
277 proteins identified in at least 3 out of 4 replicates of at least one treatment were retained for the
278 analyses. The data was background corrected and normalized using the variance stabilizing
279 transformation. Intensity distributions for proteins with and without missing values were
280 inspected to establish that the missing values were mostly related to low intensities values. Thus,
281 missing values were imputed using the “MinProb” function of the MSnbase (Gatto and Lilley,
282 2011) Bioconductor package integrated into DEP 1.5.1. The data was expressed as log₂ of the
283 fold changes (log₂-FC) compared to the control samples (DMSO). A protein was considered as
284 significantly differentially expressed if the log₂-FC $\geq |0.4|$ with the false discovery rate (FDR) for
285 the adjusted *p*-values set to 5%. The *k*-means clustering and heat map for the differentially
286 expressed proteins were produced using the DEP package with the default parameters (Euclidean
287 distance). Proteins that were identified exclusively by using the *S. salar* RefSeq database are
288 presented separately from those heatmaps. GO (Gene Ontology) terms obtained from the Uniprot
289 and QuickGO databases were used to help identifying the functions of all differentially
290 expressed proteins. As more GO annotations were available for *S. salar* than for *O. mykiss*, all *O.*
291 *mykiss* proteins Uniprot IDs were converted to *S. salar* Uniprot IDs using the NCBI BLAST
292 software 2.7.1 algorithm (National Center for Biotechnology Information, Bethesda, MD, USA).

293 Two differentially expressed proteins were labelled as “uncharacterized protein” in the *O. mykiss*
294 and *S. salar* databases. Their RefSeq sequences were submitted to BLASTp for identification.
295 The R package DEP 1.5.1 was applied for the statistical analyses of the metabolomics data as
296 well. A total of 69 annotated metabolites were included in the analysis. The data was expressed
297 as log₂ of the fold changes (log₂-FC) compared to the control samples (DMSO). For all other
298 data, the normal distribution was assessed with the Shapiro-Wilk test. For the mortality data,
299 which was expressed as proportions (%) of individuals, differences among treatments were
300 assessed using the Fisher’s exact test (FE). For other data, differences among treatments were
301 tested using a one-way ANOVA followed by multiple comparisons Tukey’s HSD test for
302 normally distributed data, and a non-parametric Kruskal-Wallis test (KW) followed by
303 Conover’s test in other cases. All analyses were performed using R 3.5.1 (The R Foundation for
304 Statistical Computing).

305 2.6.2. Pathway enrichment analysis

306 Pathways enrichment analyses (PEAs) were performed using the web-based tool PaintOmics 3
307 (v0.4.5). PaintOmics allow users to perform PEA using multiple OMICs datasets (Hernández-de-
308 Diego et al., 2018). The tool first evaluates for each dataset the subset of genes, proteins or
309 metabolites that participate in a particular KEGG pathway (Kyoto Encyclopedia of Genes and
310 Genomes, Kyoto University, Japan). It evaluates the fraction of those biological features which
311 overlaps with the set of features that the researcher considered relevant, and finally computes the
312 significance of the overlap using the Fisher’s exact test. When several tests for different datasets
313 are performed for a single pathway (e.g. for proteomics and metabolomics), a single *p*-value is
314 computed using the Fisher’s combined probability test.

315 PEAs were performed for each compound and each sampling point: only proteomics datasets
316 were included for the analyses related to day 7, while the analyses related to day 14 included
317 both the proteomics and metabolomics datasets. Proteins Uniprot IDs had to be converted to
318 zebrafish (*Danio rerio*) Entrez Gene ID prior to the PEA, since this is the only fish species and
319 gene/protein nomenclature supported by KEGG pathways. This was done by gathering each
320 RefSeq nucleotide ID linked to each Uniprot ID for *S. salar* or *O. mykiss* on the Uniprot database
321 and submitting those RefSeq nucleotide IDs to the NCBI BLAST software 2.7.1 algorithm. The
322 BLAST output was filtered to keep only the matches with E -values $\leq 10^{-3}$. Matches with E -
323 values superior to this threshold were checked manually.

324 The proteomics datasets were submitted to PaintOmics as two separate files for each compound
325 and time point: one “data” file containing the list of all proteins and their corresponding \log_2 -FC
326 and one “relevant” file containing only the list of proteins that were found to be significantly
327 differentially expressed for that particular compound and time point. Similar files were generated
328 and submitted to PaintOmics for the metabolomics datasets. All metabolites with a \log_2 -FC \geq
329 $|0.4|$ were considered as “relevant” for the PEAs.

330 **3. Results**

331 3.1. Mortality and deformities

332 The mortality and deformity data for the larvae used for the proteomics assay was already
333 presented (submitted manuscript; Supplementary file 1): no significant effects were detected
334 compared to the DMSO control group. In the larvae used for the metabolomics assay, none of the
335 PAHs significantly increased the mortality after 14 days of exposure (FE, $p > 0.05$, data not
336 shown). The highest observed mortality was equal to 9.44% (5.98-14.92) in the group of larvae

337 exposed to retene (% mortality and 95% confidence interval). The mortality of the control group
338 (exposed to DMSO only) was 4.44% (2.27-8.52). None of the tested PAHs had a significant
339 effect on the BSD severity index (KW, $p > 0.05$). This index ranged from 0.06 ± 0.05 (control
340 group) to 0.15 ± 0.13 (retene) across all treatments (mean \pm SD).

341 3.2. Water parameters and PAH concentrations

342 The characteristics of the lake water used in the metabolomics assay were as follows:
343 conductivity $26.2 \pm 1.0 \mu\text{S}$, pH 7.24 ± 0.13 , temperature $11.5 \pm 0.3^\circ\text{C}$ and oxygen saturation
344 $>96\%$. The water parameters for the proteomics assay experiments have already been presented
345 (submitted manuscript; Supplementary file 1) and were very similar to those presented above.
346 Similarly, the concentrations of each PAH in water at different time points in the proteomics
347 assay experiments have been presented separately (submitted manuscript; Supplementary file 1).
348 Due to some water samples storage issues, it was only possible to measure the concentrations of
349 pyrene in the metabolomics assay experiment. The concentration of pyrene in the tanks left
350 without fish was $10.45 \pm 2.61 \mu\text{g.L}^{-1}$ (Fig. 1A), close to what was measured in the proteomics
351 assay experiments, i.e. $8.84 \pm 1.99 \mu\text{g.L}^{-1}$ (submitted manuscript; Supplementary file 1). As
352 observed previously in the proteomics assay experiments (submitted manuscript; Supplementary
353 file 1), the concentration of pyrene measured in the water in the presence of larvae tended to
354 decrease over time (Fig. 1A). The average fluorescence spectrums for pyrene in the exposure
355 water are displayed in figure 1B. No metabolites were detected in the tanks without larvae, while
356 one peak of metabolite was detected (around 340 nm) already after 1 day of exposure in the
357 presence of larvae. This peak of metabolite was noticeably higher after 3, 7 and 10 days of
358 exposure, and then further increased after 14 days. Unlike in the proteomics assay experiments

359 (submitted manuscript; Supplementary file 1), it was unclear if another peak could be observed
360 around 360 nm.

361 3.3. Proteomics analysis output

362 A total of 28 proteins were differentially expressed in the hearts of rainbow trout larvae after 7
363 days of exposure to retene, pyrene or phenanthrene (Fig. 2). The raw output from the DEP
364 package analysis for those proteins, including all *p*-values and log₂ fold changes, is available in
365 the Supplementary file 2. Some proteins such as cytochrome P450 and sulfotransferase were
366 identified twice in the RefSeq database, most likely because different isoforms of the same
367 proteins were differentially expressed by one or several of the tested PAHs. The cytochrome
368 P450 with the highest fold change values was the only protein that was significantly
369 differentially expressed by all PAHs, while the uncharacterized protein displayed in figure 2 was
370 significantly affected by both retene and pyrene. This uncharacterized protein showed similarity
371 with the MAdCAM-1 protein (mucosal vascular addressin cell adhesion molecule 1, also known
372 as addressin), according to BLASTp. All other proteins were significantly differentially
373 expressed by only one PAH, despite some relatively high measured fold changes like in the case
374 of the other cytochrome P450 isoform following exposure to pyrene, for example. High variability
375 between replicates can explain those observations. In addition to those 28 proteins, one
376 differentially expressed protein was found exclusively when using the *S. salar* database during
377 the bioinformatics analyses: DNA (cytosine-5)-methyltransferase was significantly upregulated
378 by phenanthrene only (log₂-FC = 0.74, Supplementary file 2).

379 A total of 43 proteins were differentially expressed in the hearts of rainbow trout larvae after 14
380 days of exposure to retene, pyrene or phenanthrene (Fig. 3), in addition to 11 proteins identified
381 only when using the *S. salar* database during the bioinformatics analyses (Fig. 4). The raw output

382 from the DEP package analysis for those proteins, including all p -values and \log_2 fold changes,
383 is available in the Supplementary file 2. Only three proteins were differentially expressed by all
384 compounds after 14 days of exposure: the cytochrome P450 isoform with the highest fold change
385 values, UDP-glucuronosyltransferase-like (Fig. 3) and the protein TFG isoform X4 (Fig. 4). One
386 of the sulfotransferase isoforms as well as the UTP—glucose-1-phosphate uridylyltransferase
387 were significantly upregulated by both retene and phenanthrene (Fig. 3). Those two compounds
388 also had in common the significant depletion of several proteins: smoothelin-like protein 2, ER
389 membrane protein complex subunit 2-like, high mobility group protein B1, fatty acid-binding
390 protein (heart-like) and 60S ribosomal protein L22-like isoform X2 (Fig. 3 and 4). The protein
391 tubulin-folding cofactor B was significantly depleted by both pyrene and retene (Fig. 3). All
392 other proteins were significantly differentially expressed by only one PAH. The BLASTp search
393 suggested that the uncharacterized protein displayed in figure 3 is similar to the Rapunzel
394 protein. Overall, only 4 proteins were found to be differentially expressed after both 7 and 14
395 days of exposure: cytochrome P450 (both isoforms), hemopexin, sulfotransferase (both isoforms)
396 and UDP-glucuronosyltransferase-like. Since the hemopexin protein was identified in the *O.*
397 *mykiss* database only after 7 days and in the *S. salar* database only after 14 days, it is possible
398 that those are two different isoforms.

399 3.4. Metabolomics analysis output

400 A total of 69 identified metabolites were measured and included in the analyses. None of the
401 metabolites were significantly enriched or depleted in the hearts of rainbow trout larvae exposed
402 to retene for 14 days (Fig. 5). Significantly enriched or depleted metabolites were only detected
403 following exposure to pyrene or phenanthrene, and only octadecane-1-ol (stearyl alcohol) was
404 significantly depleted by both PAHs. Overall, a very high variability was observed in the data

405 among replicates for several metabolites. This can easily be seen in the volcano plots (Fig. 5) as
406 several metabolites had a relatively high (or low) \log_2 fold change value but were still not
407 significantly different when compared to the DMSO control group (with low $-\log_{10}$ adjusted p -
408 values) following the LIMMA analysis. Finally, it is worth mentioning that phenanthrene itself
409 was detected in each of the four replicate samples of the hearts of the rainbow trout larvae
410 exposed to phenanthrene, but not in any other exposure group (controls, retene or pyrene).
411 Retene and pyrene were not detected in any sample. The raw output from the DEP package
412 analysis for all metabolites, including all p -values and \log_2 fold changes, is available in the
413 Supplementary file 3.

414 3.5. Pathway enrichment analysis

415 Significantly enriched KEGG pathways are presented in the Table 1. The only pathways
416 common to every tested compound as well as both sampling points were related to metabolism
417 of xenobiotics, retinol metabolism and steroid hormone biosynthesis, and were linked mostly to
418 two proteins: cytochrome P450 and UDP-glucuronosyltransferase. Overall, those two proteins
419 were present in the majority of the enriched pathways. A few interesting pathways were included
420 in the presented data despite only having p -values close to the significance level. Several
421 complete KEGG pathways of interest are available in Supplementary files 4 and 5.

422 Very few relevant metabolites were involved in the presented pathways. Arabitol was the only
423 significantly altered metabolite (Fig. 5) present in an enriched pathway, i.e. the pentose and
424 glucuronate interconversions pathway in the case of phenanthrene. Phenylalanine was present in
425 two pathways altered by phenanthrene (dre00360 and dre00400), but was not significant itself in
426 the metabolite dataset (Fig. 5) despite a relatively high fold change value (\log_2 -FC = 0.55)
427 (Supplementary file 5E).

428 **4. Discussion**

429 4.1. Effects of PAHs on larvae development and concentrations of PAHs in water

430 None of the three model PAHs significantly increased the mortality or the BSD index in the
431 rainbow trout larvae used for the metabolomics assay. This is very consistent with the results of
432 the transcriptomics and proteomics assays (Supplementary file 1). The concentration of pyrene
433 over time in the metabolomics assay exposure followed a similar trend as that reported for the
434 transcriptomics assay, and its concentration in the tank without larvae was close to the one
435 previously reported (submitted manuscript; Supplementary file 1). However, several differences
436 were observed in the magnitude of the peaks of the first metabolite appearing at 342 nm
437 (speculated as being either 1-hydroxypyrene glucuronide or 1-hydroxypyrene) between the two
438 studies, especially at 7 and 14 days (submitted manuscript; Supplementary file 1). These may be
439 a result of subtle biological differences between larvae, as the ones used for the transcriptomics
440 assay and the metabolomics assays were from two different batches and used in two distinct
441 experiments. Experiments involving the quantification of pyrene and its metabolites in both
442 larvae and exposure water using a more sensitive HPLC method are currently underway.

443 4.2. PAHs cause distinct changes on the cardiac proteome

444 Overall, very few proteins were differentially expressed by more than one PAH, suggesting that
445 each PAH caused unique changes in the cardiac proteome of the rainbow trout ELS. The proteins
446 that were altered by all three compounds were mostly related to the metabolism of xenobiotics,
447 i.e. cytochrome P450 and UDP-glucuronosyltransferase. It was well translated into the PEAs as
448 well, with those two proteins being the only ones involved in metabolic pathways common to all
449 three PAHs (Table 1). This also suggests that each PAH was able to activate the AhR. These

450 phase I and II metabolism proteins as well as sulfotransferase and hemopexin were also the only
451 proteins altered at both sampling time points. Retene was the most potent cytochrome P450 and
452 UDP-glucuronosyltransferase inducer, as well as the only compound to induce another
453 cytochrome P450 isoform at significant levels. Phenanthrene had no significant effect on UDP-
454 glucuronosyltransferase after 7 days but was able to induce it after 14 days. Retene was also the
455 only compound to induce sulfotransferase, with the exception of phenanthrene also inducing one
456 of the two identified isoforms after 14 days. Unique signatures from each PAH were also
457 observed in the cardiac transcriptome, as well as similar expression patterns for the genes
458 encoding for the aforementioned proteins involved in the metabolism of xenobiotics (submitted
459 manuscript). Unique proteomic fingerprints were already described *in vitro* between different
460 PAHs or PAHs mixtures (Hooven and Baird, 2008). Pyrene and phenanthrene were equipotent to
461 significantly induce the protein level of the first isoform of cytochrome P450 after 14 days of
462 exposure (Fig. 3 and Supplementary file 2), but otherwise showed a different proteomic
463 fingerprint. This suggest that for weak AhR agonists such as these two compounds, their toxicity
464 cannot be predicted based on their ability to activate the AhR pathway.

465 High conservation between responses at the gene or protein expression levels has already been
466 reported before for various model AhR agonists in another fish species (whole body tissue), the
467 white sturgeon (*Acipenser transmontanus*) (Doering et al., 2016). At the transcriptomic level,
468 retene altered the expression of several genes involved in the generation of the cardiac action
469 potential, ion homeostasis (including calcium) and the sarcomere (actin, myosin and troponin) at
470 different time points of the early heart development (submitted manuscript). However, as most of
471 these changes at the gene expression level were observed much earlier during development (i.e.
472 after 1 and 3 days of exposure), it is possible that the present proteomics assay conducted after 7

473 and 14 days of exposure missed most of the relevant changes at the protein level. A similar
474 remark can be made for pyrene, which was found to alter several myosin-related genes after 1
475 day of exposure (submitted manuscript). Performing proteomics on the cardiac tissue of rainbow
476 trout larvae exposed to those two PAHs for 1 or 3 days is technically feasible, but would require
477 more time, skilled manpower and resources, as it would require more than 45 individual hearts
478 being dissected and pooled per sample.

479 Nonetheless, concomitant results can be highlighted from the proteomics data from the present
480 study. Retene depleted proteins such as calreticulin after 7 days, as well as alpha-actinin-1 (not
481 significant), tropomyosin alpha-4 and myosin-11 after 14 days. Interestingly, calreticulin is a
482 protein of similar function as calsequestrin, as both are involved in the storage of calcium in the
483 endoplasmic reticulum for the former and in the sarcoplasmic reticulum for the latter (Lee and
484 Michalak, 2010). Retene was already downregulating the expression of at least one gene related
485 to calsequestrin (*casq1b*) in the rainbow trout developing heart (submitted manuscript). Calcium
486 homeostasis is a known target of PAHs mixture in fish (Greer, J. B. et al., 2019; Xu et al., 2016;
487 Xu et al., 2019). The present study suggests that retene is also able to alter calcium storage and
488 homeostasis by reducing the calreticulin content at the protein level. While it is calsequestrin and
489 not calreticulin that is known to play a direct role in the control of excitation-contraction
490 coupling of cardiomyocytes, recent studies in mice suggests that calreticulin-dependent
491 homeostasis of calcium is also important to maintain the normal physiological function of the
492 heart. Calreticulin may indirectly affect the calcium cycling-proteins and gap junctions of the
493 sarcoplasmic reticulum (Lee et al., 2013). Another interesting protein affected by retene in
494 regards to calcium handling is the peptidylpropyl isomerase, which was significantly depleted
495 after 14 days of exposure (Fig. 4). The gene encoding for the peptidylpropyl isomerase 1 (*Pin1*)

496 was shown to play a role in calcium handling in mice cardiomyocytes, by influencing the
497 SERCA (sarco/endoplasmic reticulum Ca^{2+} -ATPase) pump and the NCX1 (sodium/calcium)
498 exchanger protein levels (Sacchi et al., 2017). The alteration of proteins related to calcium
499 handling and myocardium architecture (including one tropomyosin isoform) was already
500 reported in the adult heart of zebrafish exposed to the highly potent AhR agonist TCDD (Zhang,
501 J. et al., 2013).

502 Two very important KEGG pathways in regards to the cardiac function appeared to be altered by
503 retene: vascular smooth muscle contraction (although only close to be significant, $p = 0.07$,
504 Supplementary file 4A) at day 7 and tight junctions at day 14 (Supplementary file 5C). Several
505 proteins related to myosins appeared to be depleted in both pathways (although not always
506 significantly). Again, comparable results were found in our transcriptomic assay, as several
507 genes related to myosins or claudins (key components of tight junctions) were significantly
508 altered by retene (submitted manuscript). The only protein significantly altered in the case of the
509 vascular smooth muscle contraction pathway was one isoform of the Rho guanine nucleotide
510 exchange factor 1. Guanine nucleotide exchange factors (GEFs) are known to play key roles in
511 the angiogenesis and the regulation of the vascular function (Kather and Kroll, 2013). Another
512 notable protein highly depleted by retene was similar to the junctional adhesion molecule C.
513 Involved in tight junctions, junctional adhesion molecules play important roles in the
514 angiogenesis and the regulation of vascular permeability in mammals (Weber et al., 2007).

515 Retene is known to induce oxidative stress and DNA damage (Maria et al., 2005; Peixoto et al.,
516 2019). In the present study, retene was the only PAH to significantly induce the superoxide
517 dismutase protein in the cardiac tissue, probably as a response to an increased content in reactive
518 oxygen species (ROS). Retene also depleted several proteins involved in DNA damage repair

519 (Supplementary file 5A) and altered the level of several proteins in the FoxO signaling pathway
520 (Supplementary file 5B), including the 5'-AMP-activated protein kinase subunit beta 1 (\log_2 -FC
521 = -1.20) related to the AMPK enzyme. In mammalian cells, the FoxO signaling pathway is
522 known to play a role in DNA damage repair, especially via the FOXO3 transcription factor
523 (Bigarella et al., 2017; Tran et al., 2002). AMPK has also been shown to interact with FOXO3 in
524 mechanisms of defense against oxidative stress (Greer, E. L. et al., 2007; Li et al., 2009).
525 Although our data may suggest that retene could promote DNA damage not only by inducing
526 oxidative stress but also by altering the AMPK enzyme and other proteins involved in DNA
527 repair, more research is needed to explore that hypothesis.

528 Both retene and pyrene significantly reduced the protein level of tubulin-folding cofactor B (Fig.
529 3). Tubulin-folding cofactors control the availability of tubulin subunits in eukaryotic cells
530 (Szymanski, 2002). Retene also slightly but significantly reduced the protein level of tubulin beta
531 chain (also known as beta tubulin) (Fig. 3). The present study is not the first to report such
532 findings for PAHs. Tubulins alpha and beta protein levels were altered *in vitro* following
533 exposure to benzo[a]pyrene, dibenzo[a,l]pyrene or coal tar extract (Hooven and Baird, 2008).
534 Interestingly, beta tubulin has been shown to play a role in the AhR function in mammalian cells:
535 it could reduce the binding of the AhR/Arnt (Ah receptor nuclear translocator) heterodimer to the
536 DRE (dioxin response element) by interacting with Arnt but not the AhR (Zhang, T. et al., 2010).
537 One can hypothesize that a reduced beta tubulin biosynthesis or protein level could result in the
538 opposite effect, i.e. an increased binding of the heterodimer with the DRE. Additional research in
539 fish regarding the role of beta tubulin is needed to explore that hypothesis, as well as its relative
540 importance in the regulation of the AhR-mediated toxicity of PAHs.

541 At the transcriptomic level, pyrene disrupted the expression of multiple genes involved in the
542 myosin complex as early as after 24 hours of exposure, and in an opposite way compared to
543 retene (submitted manuscript). In the present work, pyrene depleted alpha-actinin-1 after 14 days
544 of exposure (Fig. 3), as well as myosin-11, but the trend was not significant for the latter. In
545 vertebrates, alpha-actinin 2 and 3 are the two muscular forms of actinin, while alpha-actinin 1
546 and 4 are more broadly expressed (Holterhoff et al., 2009). Those non-muscle forms of alpha-
547 actinin not only bind with actin, but also play a role in stress fibers, focal adhesions, the
548 cytoskeleton, as well as in adherens and tight junctions (Otey and Carpen, 2004). The adherens
549 junction KEGG pathway was indeed close to being significantly enriched following the PEAs (p
550 =0.06, Table 1). This pathway involved not only alpha-actinin-1, but also a protein similar to
551 catenin delta-1, which had a relatively highly negative log₂-FC value (-0.97, data not shown) but
552 was not significantly depleted. The alteration of proteins related to myosin, actin, tropomyosin or
553 more generally to the muscular system development and function was already reported for fish
554 exposed to PAHs mixtures (Bohne-Kjersem et al., 2010; Karlsen et al., 2011; Simmons and
555 Sherry, 2016). Our data suggest that both retene and pyrene alone are able to alter such proteins.
556 Early exposure to these PAHs could result in a loss in cardiac fitness later in the life of the
557 juvenile fish, as demonstrated with PAHs mixture in several fish species (Hicken et al., 2011;
558 Incardona et al., 2015).

559 The most interesting changes in protein expression linked to the pyrene exposure were related to
560 the metabolism and handling of iron. Iron-mediated oxidation-reduction reactions are required
561 for the proper metabolism of oxygen in the heart. Iron levels in the heart are tightly regulated,
562 and consequently heart failure is a common denominator in conditions of systemic iron
563 imbalance (Lakhal-Littleton, 2019). In the present study, pyrene significantly depleted

564 hemopexin from the cardiac tissue of the exposed rainbow trout larvae, especially after 7 days of
565 exposure (Fig. 2). Hemopexin is a plasma protein with a very high binding affinity for heme
566 complexes (cytochromes, iron and porphyrin complexes), and it plays a protective role against
567 heme-induced oxidative stress (Ingoglia et al., 2017; Tolosano and Altruda, 2002). Altered gene
568 expression or protein levels for hemopexin were already reported in several studies involving
569 different fish species exposed to benzo[a]pyrene or complex PAHs mixtures (Alderman et al.,
570 2017; Bohne-Kjersem et al., 2009; Enerstvedt et al., 2017; Karlsen et al., 2011; Won et al.,
571 2013). Interestingly, pyrene was also able to increase the expression of several proteins involved
572 in porphyrin metabolism (Supplementary file 4B), notably uroporphyrinogen decarboxylase
573 (UROD) which was also upregulated at the transcriptomic level after 7 days of exposure in our
574 related study (submitted manuscript). Finally, several proteins involved in the ferroptosis
575 pathway (Supplementary file 4D) and most notably ferritin were also enriched following
576 exposure to pyrene, suggesting a response to an increase in free iron ion content. It is worth
577 mentioning that pyrene also upregulated two genes related to the hemoglobin complex, *hbae1.3*
578 and *hbae3*, as well as two genes linked to ferritin and transferrin, in the aforementioned
579 transcriptomic study. In fish, PAHs mixtures have been previously shown to alter the protein
580 levels of hemoglobin and other proteins involved in iron metabolism (Pampanin et al., 2014).
581 Phenomenon such as hemolysis (destruction of red blood cells) are known to release hemoglobin
582 and free heme into the circulation. Free toxic hemes are generally scavenged by hemopexin and
583 subsequently catabolized into carbon monoxide (CO), biliverdin, and ferrous iron (Fe^{2+}) by heme
584 oxygenase-1 (HO-1). In severe hemolysis, the extensive release of heme from hemoglobin or
585 decreased hemopexin levels lead to an increase in the iron level in the circulation, which is
586 consequently handled by ferritin (Chiang et al., 2019; NaveenKumar et al., 2018). In the present

587 study, pyrene clearly appeared to act as a disruptor of heme and iron metabolism in the rainbow
588 trout ELS heart, although the exact triggering mechanisms remains to be found. In birds exposed
589 to PAHs, metabolism of parent PAHs by CYP1A creates oxidative PAH metabolites that cause
590 oxidative damage to erythrocytes, resulting in hemolytic anemia associated with an increase in
591 ferritin (Troisi et al., 2007). Hemolytic anemia following exposure to naphthalene or complex
592 PAHs mixtures has also been reported *in vitro* for mammals (Couillard and Leighton, 1993).
593 Interestingly, genes coding for HO-1 and hemopexin are AhR-regulated genes, and an *in vitro*
594 exposure of human liver cells to benzo[a]pyrene resulted in a decreased hemopexin gene
595 expression (Iwano et al., 2010).

596 Phenanthrene appeared to be the least potent of all three tested PAHs in regards to changes in the
597 cardiac proteome after 7 days. Besides the significant induction of cytochrome P450, one of the
598 most notable change was the depletion of both alpha and beta fibrinogen chains (Fig. 2).
599 Fibrinogen is important to heal tissue and blood vessels injuries, as it forms fibrin-based blood
600 clots following enzymatic conversion by thrombin. A reduced fibrinogen content in larvae
601 exposed to phenanthrene could indicate a reduced ability to recover from such damage, or a
602 response to tissue and vascular damage potentially caused by phenanthrene. Proteins involved in
603 the fibrinolytical system (including fibrinogen) were altered in the juvenile cod (*Gadus morhua*)
604 exposed to a complex PAHs mixture (Bohne-Kjersem et al., 2009). After 14 days, phenanthrene
605 had in common with retene the significant depletion of smoothelin (Fig. 3), a protein found in
606 vascular smooth muscles that has been shown to interact with calmodulin and troponin (Ulke-
607 Lemée et al., 2017). In mammals, a deficiency in smoothelin has been associated with reduced
608 vascular contractility, and is often observed after vascular damages (Rensen et al., 2008; van Eys
609 et al., 2007). Another protein significantly depleted by both retene and phenanthrene after 14

610 days of exposure was the heart-like fatty acid-binding protein (Fig. 3), a protein responsible of
611 the intracellular transportation of long-chain fatty acid and found in abundance in
612 cardiomyocytes (Schaap et al., 1998). Some proteins involved in the metabolism of fatty acids
613 were also significantly altered by phenanthrene, i.e. long-chain-fatty-acid—CoA ligase 1-like
614 and aldehyde dehydrogenase family 16 member A1 (Fig. 3 and Table 1). Fatty acids are an
615 important source of energy for the heart, and our data suggest that phenanthrene might affect
616 both their transport and metabolism. Aldehyde dehydrogenase, heart-like fatty acid-binding
617 protein as well pyruvate dehydrogenase were all significantly altered by phenanthrene (Fig. 3),
618 which is consistent with the results reported in the liver proteome of the largemouth bass
619 (*Micropterus salmoides*) exposed to phenanthrene too (Sanchez et al., 2009). Pyruvate
620 dehydrogenase catalyzes the oxidative decarboxylation of pyruvate generated from glycolysis,
621 producing NADH, carbon dioxide and acetyl coenzyme A (acetyl-CoA). It is an essential link
622 between the glycolysis and the TCA cycle located in the mitochondria, and plays an important
623 role in the oxidative consumption of glucose (Tzagoloff, 2012).

624 4.3. Changes induced in the metabolites profile of the rainbow trout heart

625 No significant changes were observed for any of the measured metabolites following exposure to
626 retene (Fig. 5). Overall, a very high variability was observed in the data among replicates for
627 several metabolites and for all compounds. Increasing the sample size (N = 4) would have
628 probably helped to detect more changes in the cardiac metabolome in the present study. Even
629 though several metabolites were either significantly depleted or enriched in the hearts of pyrene
630 or phenanthrene-treated larvae, very few could be linked to the proteomics datasets and mapped
631 into any KEGG pathways following the PEAs, making the interpretation of our datasets difficult.
632 The only significant metabolite mapped to a KEGG pathway was arabitol, and was in excess

633 after exposure to phenanthrene and involved in the pentose and glucuronate interconversions
634 pathway as well (Fig. 5 and Table 1). The slight but significant valine deficiency observed for
635 the phenanthrene exposure (Fig. 5) could be indicative of a disorder in the catabolism of that
636 particular BCAA (branched-chain amino acids). Altered or defective BCAA catabolism can
637 potentially be linked to heart failure (Sun, H. and Wang, 2016). Another interesting enriched
638 metabolite in the case of phenanthrene was phenylalanine. It was associated to a significant
639 induction of the expression of one aspartate aminotransferase (also known as aspartate
640 transaminase) isoform (Supplementary file 5E), which catalyzes the transformation of
641 phenylalanine to phenylpyruvate. However, even though phenylalanine had a relatively high fold
642 change, it was not significantly enriched according to the metabolomics data analyses, making it
643 difficult to draw any conclusions about it.

644 Myo-inositol (often simply referenced as inositol) was slightly but significantly in excess in the
645 cardiac tissue of rainbow trout ELS following pyrene exposure. Inositol is an important basis for
646 several secondary messengers such as inositol trisphosphate, which for example plays important
647 roles in the calcium homeostasis of cardiomyocytes (Garcia and Boehning, 2017). More research
648 is needed to assess if pyrene can possibly affect the cardiac function by altering the inositol
649 signaling pathway. Alanine content was significantly lowered by pyrene in the cardiac tissue
650 compared to control larvae. Alanine is an important amino acid for protein biosynthesis, but is
651 also used as an alternative source of energy by being converted into pyruvate (used to produce
652 glucose) in the liver or in muscles. Low levels in alanine could thus indicate an increased energy
653 demand or an altered glucose cycle in the cardiac tissue of larvae exposed to pyrene. Pantothenic
654 acid, also known as vitamin B₅, appeared to be in slight excess following exposure to pyrene.
655 Pantothenic acid is required for the biosynthesis of coenzyme A (CoA), which is involved in

656 various key biological processes such as the oxidation of fatty acids, carbohydrates, pyruvate,
657 lactate, ketone bodies, and amino acids, as well as many other biosynthesis reactions (Tahiliani
658 and Beinlich, 1991). An excess in pantothenic acid could mean an altered CoA synthesis and
659 thus possible repercussions on all the aforementioned biological processes.

660 **4. Conclusion**

661 This study highlights some possible mechanisms of cardiotoxicity of retene and pyrene in fish
662 ELS. Retene altered several key proteins related to muscle contraction, cellular tight junctions or
663 calcium homeostasis. Those observations are very consistent with some results that we obtained
664 at the transcriptomic level (submitted manuscript), where retene was found to alter numerous
665 genes linked to key cardiac ion channels, the sarcomere or intercellular junctions. The most
666 notable finding of the present study was the ability of pyrene alone to disrupt the levels of
667 several proteins linked with the metabolism and handling of iron and heme. While such
668 mechanisms have been mentioned already in the literature for fish or birds exposed to PAHs
669 mixtures or benzo[a]pyrene, our data suggests that pyrene, one of the most widespread PAH in
670 the environment, could be an important contributor to this toxic pathway. More research is
671 needed to confirm that hypothesis, as well as to determine if those changes are happening
672 directly (i.e. activation of the AhR by pyrene) or indirectly (i.e. for example via hemolysis
673 induced by oxidative stress). As observed at the transcriptomic level, no clear mechanisms or
674 pathways critically relevant to the fish ELS cardiac function could be highlighted for
675 phenanthrene.

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682 **Data accessibility statement**

683 The mass spectrometry proteomics data have been deposited to the ProteomeXchange
684 Consortium via the PRIDE partner repository with the dataset identifier PXD017294.

685 **Author contributions**

686 **Cyril Rigaud:** Conceptualization, Methodology, Data Acquisition, Data Analysis, Writing –
687 Original Draft. **Andreas Eriksson:** Conceptualization, Methodology, Data Acquisition, Data
688 Analysis, Writing – Review & Editing. **Anne Rokka:** Methodology, Data Acquisition, Writing –
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691 Methodology. **Heli Lehtivuori:** Methodology. **Eeva-Riikka Vehniäinen:** Conceptualization,
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693 Editing.

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951

952 **FIGURE AND TABLE CAPTIONS**

953 **Fig. 1. (A)** Boxplots of the concentrations of pyrene measured in water by SFS after 1, 3, 7, 10 or
954 14 days of exposure (N = 3) or in tanks without fish (N = 14). **(B)** Average fluorescence
955 spectrums for pyrene in the exposure water after 1, 3, 7, 10 or 14 days (N = 3), and in tanks
956 without fish (N = 14).

957 **Fig. 2.** Heatmap displaying the proteins that were found to be differentially expressed in the
958 heart of rainbow trout larvae, compared to the control group (DMSO) and after 7 days of
959 exposure to either retene (RET, 32 $\mu\text{g.L}^{-1}$), pyrene (PYR, 32 $\mu\text{g.L}^{-1}$) or phenanthrene (PHE, 100
960 $\mu\text{g.L}^{-1}$). Proteins are displayed in rows and grouped according to similarities between treatments

961 by *k*-means clustering. Fold changes compared to the control group are expressed as \log_2 values
962 and the color scale indicates the intensity. The proteins displayed in the present figure were
963 identified using the *O. mykiss* RefSeq database. N = 4 replicates per treatment and protein. *
964 indicates a significant difference compared to the control group.

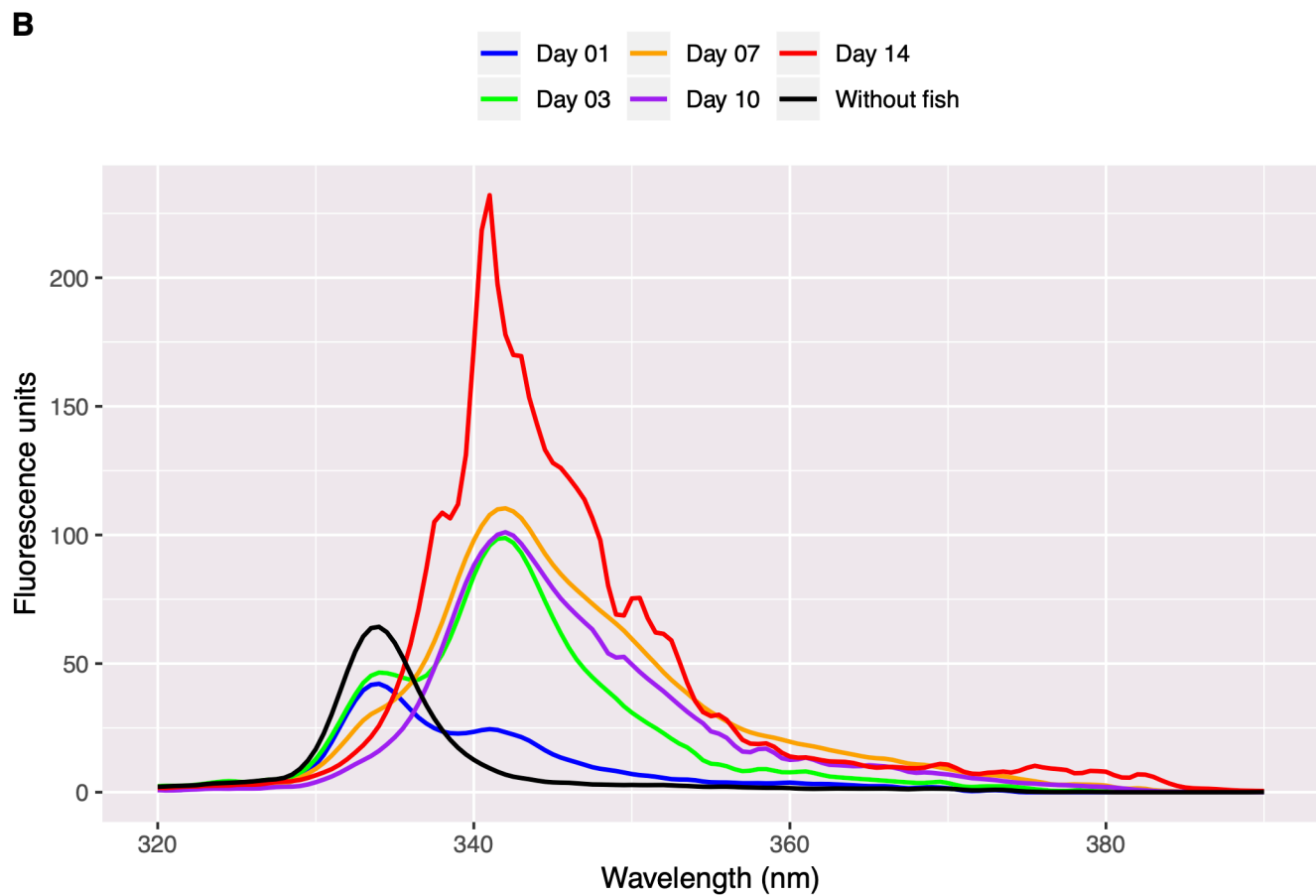
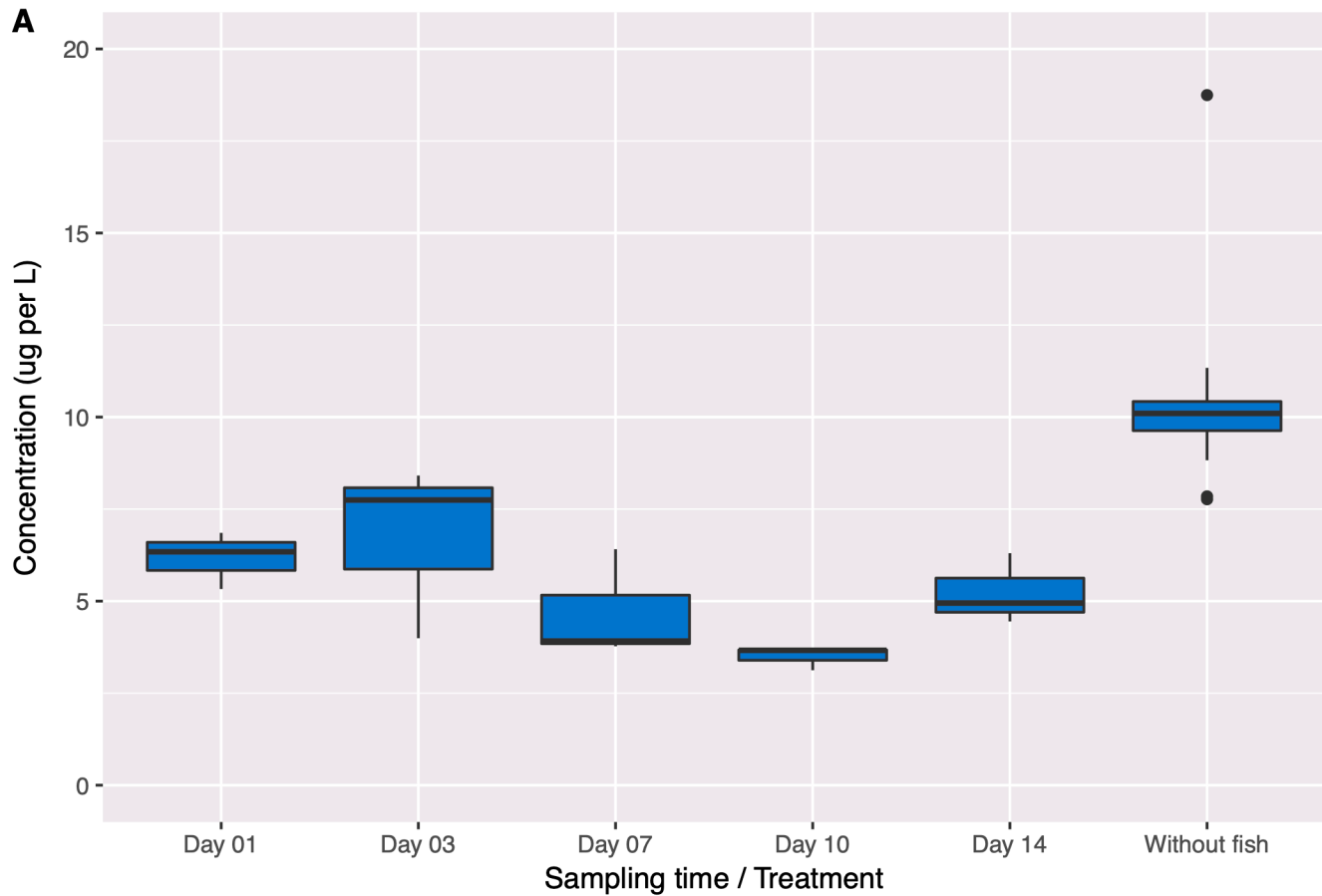
965 **Fig. 3.** Heatmap displaying the proteins that were found to be differentially expressed in the
966 heart of rainbow trout larvae, compared to the control group (DMSO) and after 14 days of
967 exposure to either retene (RET, 32 $\mu\text{g.L}^{-1}$), pyrene (PYR, 32 $\mu\text{g.L}^{-1}$) or phenanthrene (PHE, 100
968 $\mu\text{g.L}^{-1}$). Proteins are displayed in rows and grouped according to similarities between treatments
969 by *k*-means clustering. Fold changes compared to the control group are expressed as \log_2 values
970 and the color scale indicates the intensity. The proteins displayed in the present figure were
971 identified using the *O. mykiss* RefSeq database. N = 4 replicates per treatment and protein. *
972 indicates a significant difference compared to the control group.

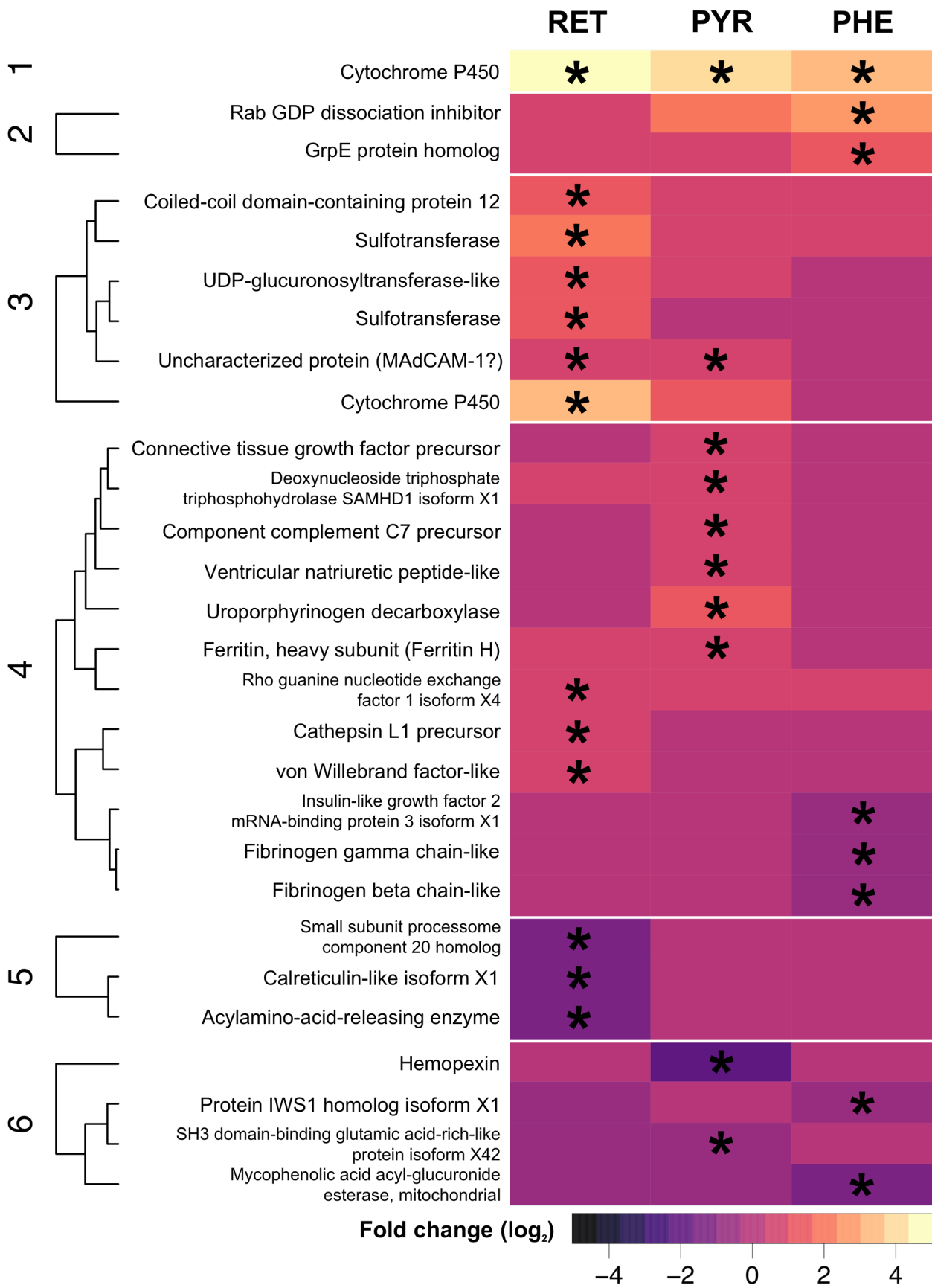
973 **Fig. 4.** Bar plot showing the proteins identified exclusively by using the *S. salar* RefSeq database
974 and that were differentially expressed compared to the control group (DMSO) after 14 days of
975 exposure to either retene (RET, 32 $\mu\text{g.L}^{-1}$), pyrene (PYR, 32 $\mu\text{g.L}^{-1}$) or phenanthrene (PHE, 100
976 $\mu\text{g.L}^{-1}$). Fold changes compared to the control group are expressed as \log_2 values and the error
977 bars represents the 95% confidence intervals. N = 4 replicates per treatment and protein. *
978 indicates a significant difference compared to the control group. ¹ Full name: SWI/SNF-related
979 matrix-associated actin-dependent regulator of chromatin subfamily D member 3-like isoform
980 X6.

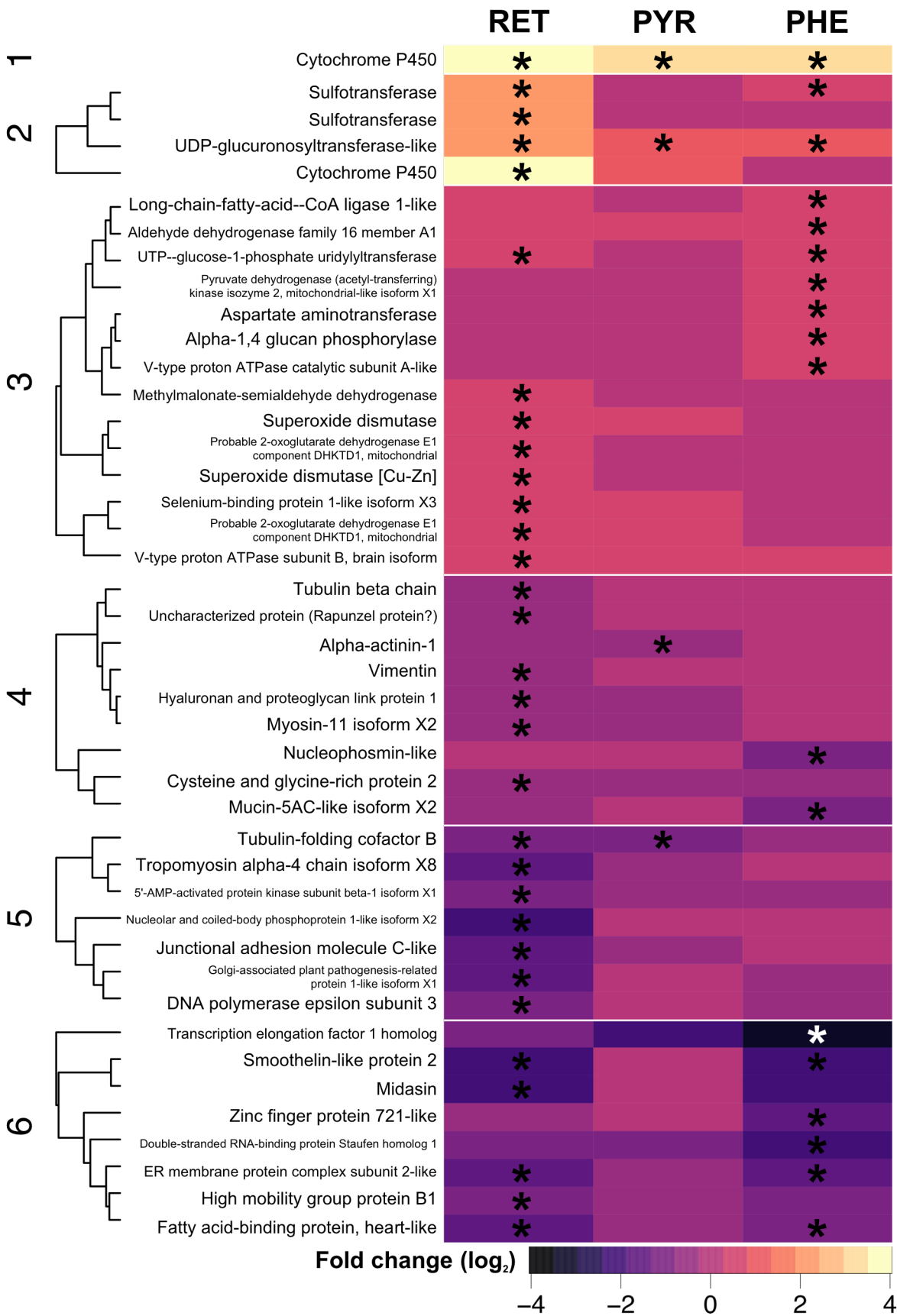
981 **Fig. 5.** Volcano plots of the metabolites identified in the heart of rainbow trout larvae after 14
982 days of exposure to either retene (RET, 32 $\mu\text{g.L}^{-1}$), pyrene (PYR, 32 $\mu\text{g.L}^{-1}$) or phenanthrene
983 (PHE, 100 $\mu\text{g.L}^{-1}$). Fold changes (x-axis) are expressed as \log_2 values compared to the control

984 group (DMSO). The y-axis displays the $-\log_{10}$ adjusted p -values. The metabolites that were
985 differentially expressed compared to the control group are featured in bold font and dark dots.
986 Greyed dots represent the metabolites that were not significantly different compared to controls.
987 $N = 4$ replicates per treatment and metabolite.

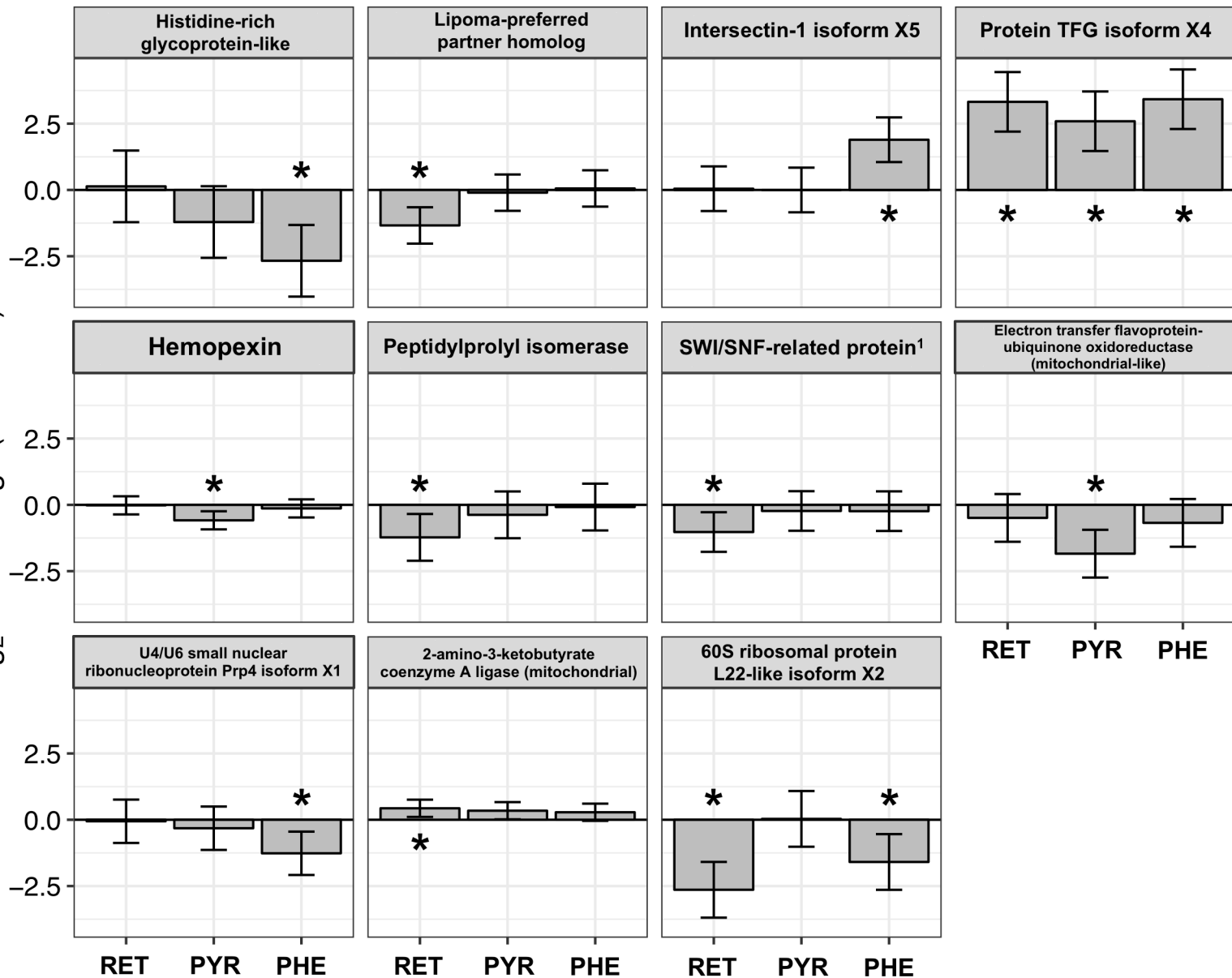
988 **Table 1.** Pathway (KEGG) enrichment analysis output from the cardiac tissue of rainbow trout
989 larvae exposed to retene (RET, $32 \mu\text{g.L}^{-1}$), pyrene (PYR, $32 \mu\text{g.L}^{-1}$) or phenanthrene (PHE, 100
990 $\mu\text{g.L}^{-1}$) for either 7 or 14 days. The number of proteins or metabolites indicate the number of
991 those that were detected and involved in the corresponding pathways, but not necessarily
992 significantly differentially expressed in the datasets.



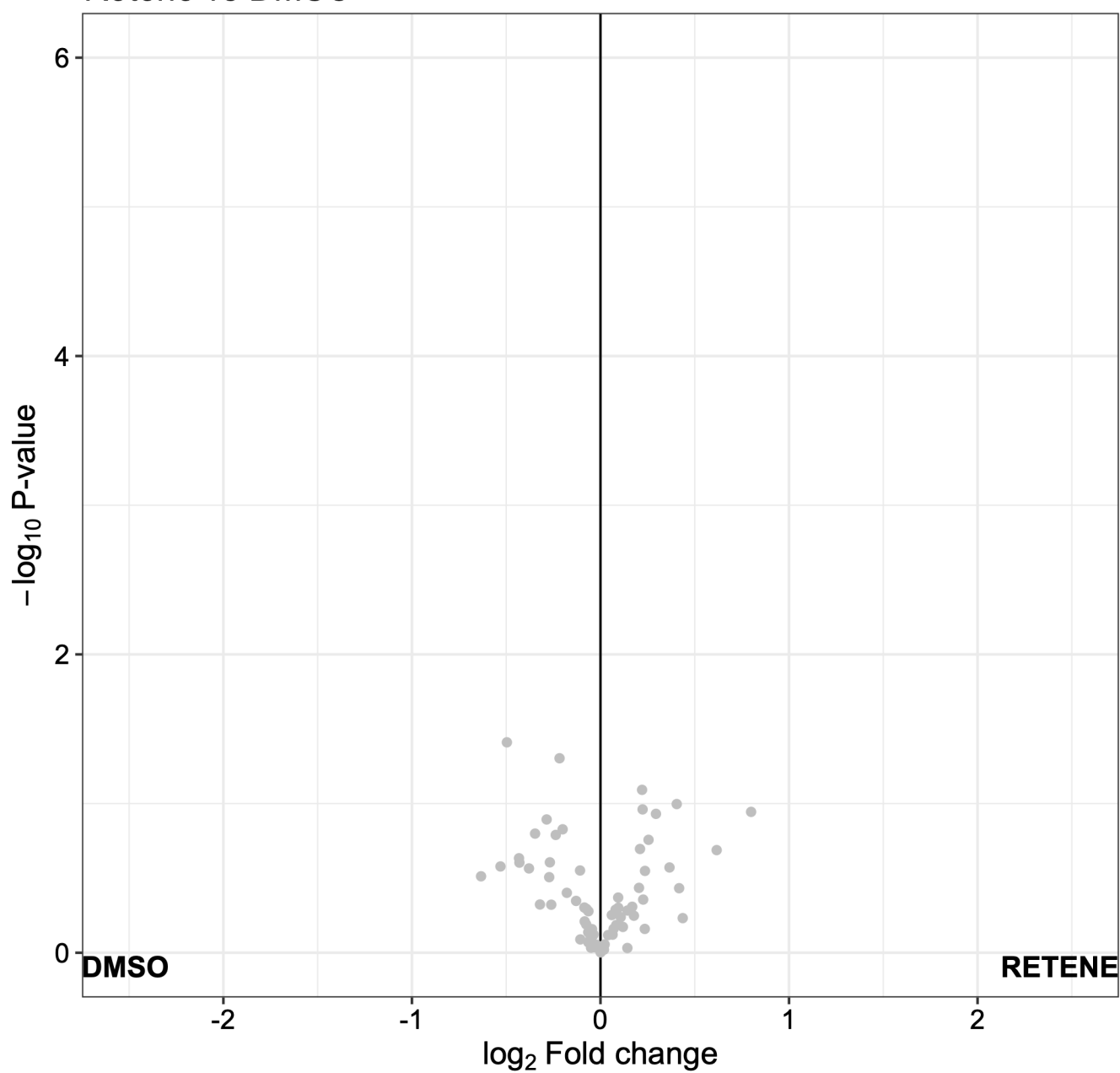




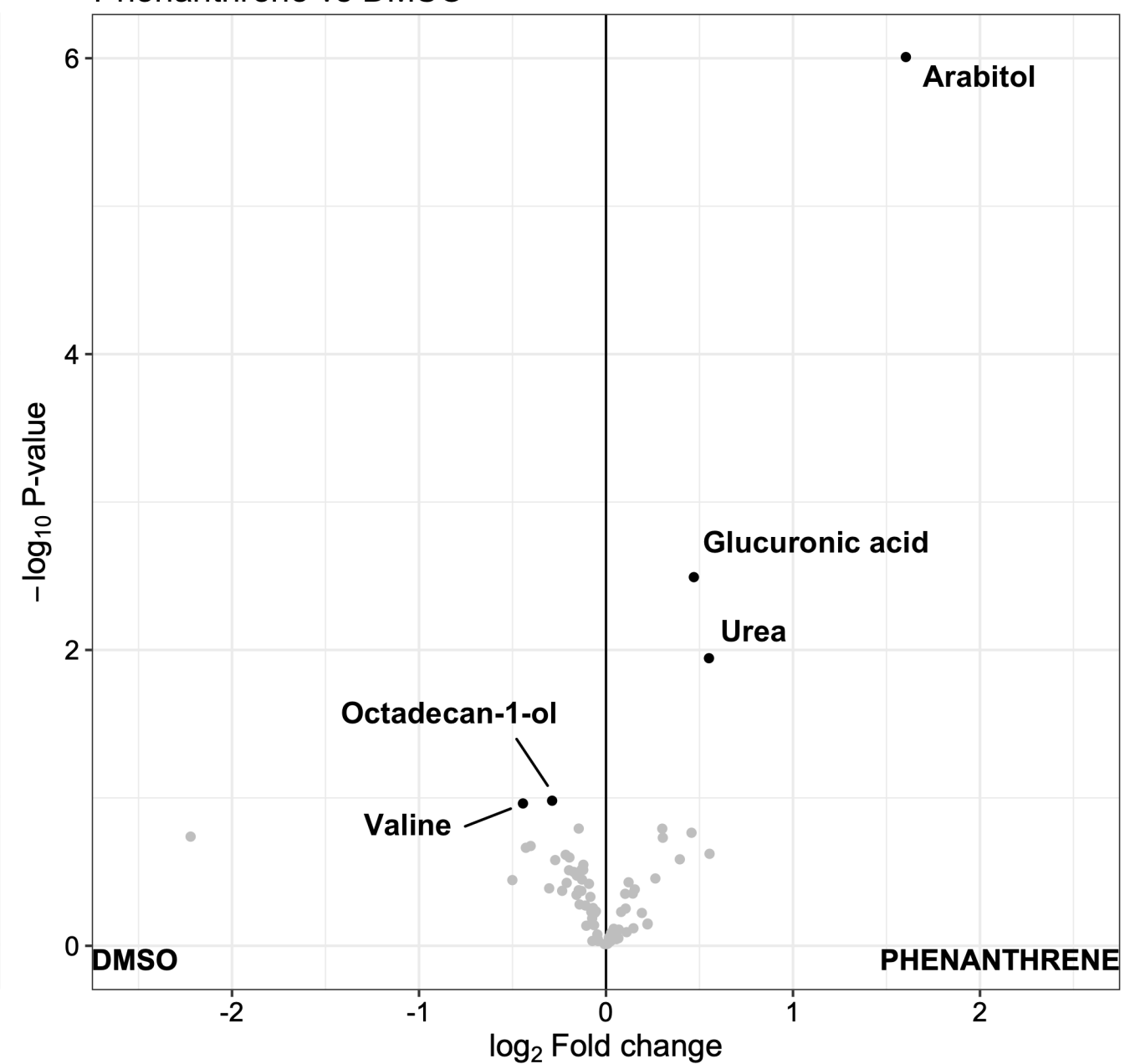
log₂ Fold change (±95% CI)



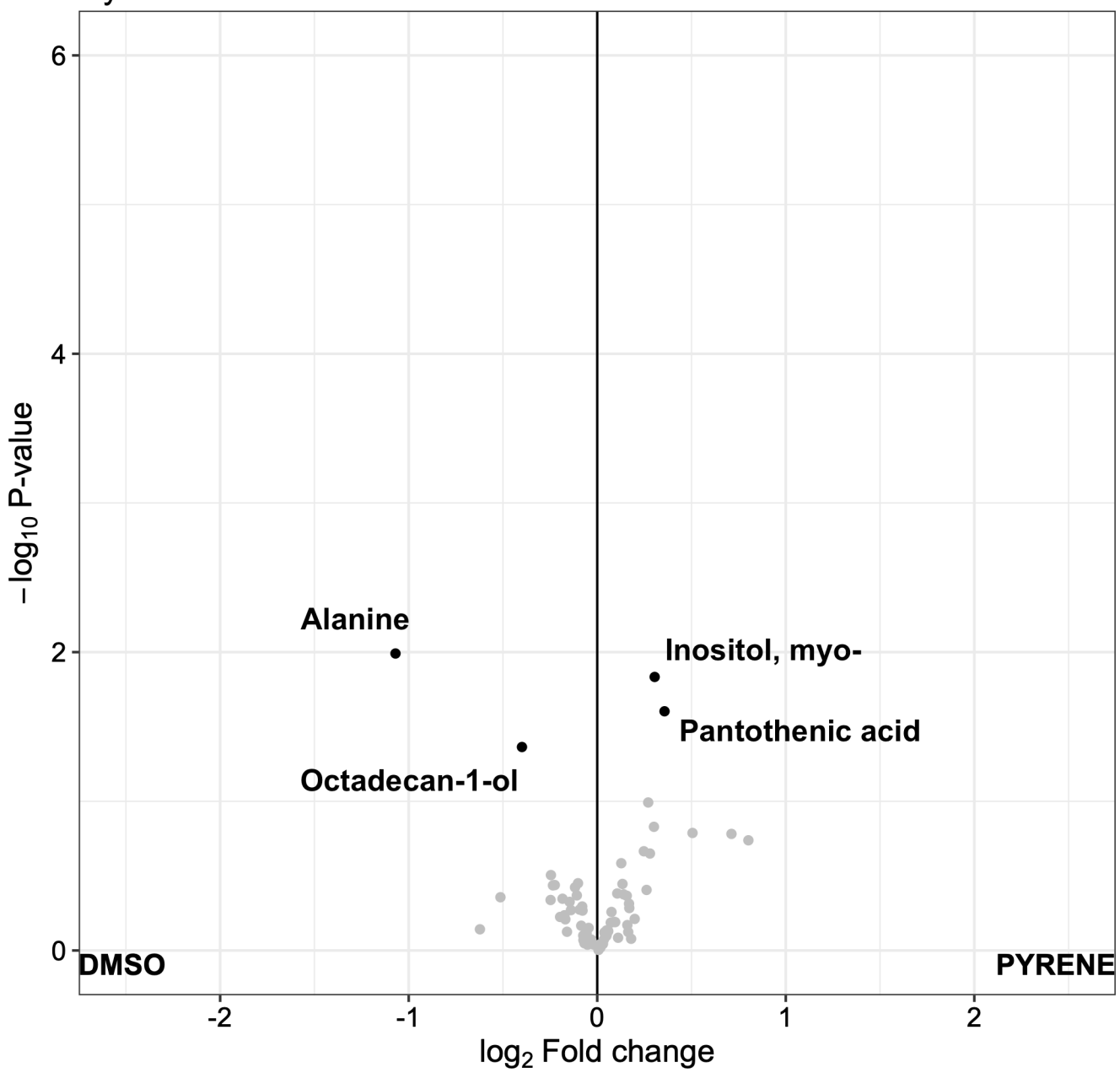
Retene vs DMSO



Phenanthrene vs DMSO



Pyrene vs DMSO



Sampling point (days)	Compound(s)	KEGG pathway	Pathway name	Proteins	Metabolites	p-value	Main proteins or metabolites involved
7	RET	dre00053	Ascorbate and aldarate metabolism	6	NA	$p \leq 0.05$	UDP-glucuronosyltransferase and UDP-glucose 6-dehydrogenase (NS ¹)
		dre00982	Drug metabolism - cytochrome P450	8	NA	$p \leq 0.05$	UDP-glucuronosyltransferase
		dre00983	Drug metabolism - other enzymes	5	NA	$p \leq 0.05$	UDP-glucuronosyltransferase
		dre00040	Pentose and glucuronate interconversions	4	NA	$p \leq 0.05$	UDP-glucuronosyltransferase
		dre04145	Phagosome	28	NA	$p \leq 0.05$	Cathepsin and calcitriculin
	RET & PYR RET, PYR & PHE PYR	dre04270	Vascular smooth muscle contraction	14	NA	$p = 0.07$	See supplementary file 2A
		dre00860	Porphyrin and chlorophyll metabolism	9	NA	$p \leq 0.05$	See supplementary file 2B
		dre00980	Metabolism of xenobiotics by cytochrome P450	11	NA	$p \leq 0.05$	Cytochrome P450 and UDP-glucuronosyltransferase
		dre00830	Retinol metabolism	3	NA	$p \leq 0.05$	Cytochrome P450 and UDP-glucuronosyltransferase
		dre00140	Steroid hormone biosynthesis	4	NA	$p \leq 0.05$	Cytochrome P450 and UDP-glucuronosyltransferase
	dre04371	Apelin signaling pathway	20	NA	$p = 0.08$	See supplementary file 2C	
	dre04216	Ferroptosis	10	NA	$p \leq 0.05$	See supplementary file 2D	
14	RET	dre03410	Base excision repair	8	0	$p \leq 0.05$	See supplementary file 3A
		dre04068	FoxO signaling pathway	13	2	$p \leq 0.05$	See supplementary file 3B
		dre00562	Inositol phosphate metabolism	3	1	$p \leq 0.05$	Methylmalonate-semialdehyde dehydrogenase
		dre04146	Peroxisome	14	0	$p \leq 0.05$	Superoxide dismutase
		dre04530	Tight junction	46	0	$p \leq 0.05$	See supplementary file 3C
	RET and PHE	dre00040	Pentose and glucuronate interconversions	4	2	$p \leq 0.05$	UDP-glucuronosyltransferase, UTP--glucose-1-phosphate uridylyltransferase, UDP-glucose 6-dehydrogenase (NS ¹) and arabinol (PHE)
		dre00500	Starch and sucrose metabolism	8	2	$p \leq 0.05$	UTP--glucose-1-phosphate uridylyltransferase, glucose (RET, NS ¹), hexokinase (RET, NS ¹) and alpha-1,4 glucan phosphorylase (PHE)
	RET, PYR & PHE	dre00980	Metabolism of xenobiotics by cytochrome P450	11	0	$p \leq 0.05$	Cytochrome P450 and UDP-glucuronosyltransferase
		dre00830	Retinol metabolism	3	0	$p \leq 0.05$	Cytochrome P450 and UDP-glucuronosyltransferase
		dre00140	Steroid hormone biosynthesis	3	1	$p \leq 0.05$	Cytochrome P450 and UDP-glucuronosyltransferase
	PYR	dre04520	Adherens junction	18	0	$p = 0.06$	Alpha-actinin-1 and catenin delta-1 (NS ¹)
		dre00982	Drug metabolism - cytochrome P450	9	0	$p \leq 0.05$	UDP-glucuronosyltransferase
		dre00983	Drug metabolism - other enzymes	5	1	$p \leq 0.05$	UDP-glucuronosyltransferase
	PYR & PHE	dre00380	Tryptophan metabolism	14	0	$p \leq 0.05$	See supplementary file 3D
	PHE	dre00053	Ascorbate and aldarate metabolism	6	2	$p \leq 0.05$	UDP-glucuronosyltransferase and aldehyde dehydrogenase
		dre00071	Fatty acid degradation	21	0	$p \leq 0.05$	Long-chain-fatty-acid--CoA ligase 1 and aldehyde dehydrogenase
		dre00360	Phenylalanine metabolism	3	1	$p \leq 0.05$	See supplementary file 3E
dre00400		Phenylalanine, tyrosine and tryptophan biosynthesis	2	1	$p \leq 0.05$	Aspartate aminotransferase and phenylalanine	

¹ Not significant.